

Economic Consequences of Childhood Exposure to Urban Environmental Toxins

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Childhood lead exposure causes well-documented damage to cognition, health, and mortality. Whether, and how, these harms translate into lasting labor market penalties is less clear. We estimate the consequences of childhood lead exposure on adult labor market outcomes by linking 6.2 million men, ages 20–50, from the 1940 full-count U.S. Census to their earliest childhood census record across 971 towns. Identification comes from the interaction of municipal pipe materials with geologically determined water chemistry that causes corrosion, leading to elevated lead leaching. Comparing across pipe types, men who were ages 0 to 10 in pure lead pipe towns with corrosive water earned 4.7% lower wages, worked 0.67 fewer weeks per year, and were 1.1 percentage points more likely to be unemployed than men in non-lead towns. The more tightly identified comparison exploits chemistry variation within pipe type: comparing across pure lead towns, men in high leaching risk water environments earned 3.6% less than those in low leaching risk environments, while the same chemistry variation has no differential effect on outcomes in towns without lead pipes. The wage penalty operates through both occupational sorting into lower-paying jobs and reduced earnings within the same detailed occupation. Lead exposure reduces childhood school attendance but does not affect self-reported completed years of education. Elevated lead exposure also compresses the relationship between childhood socioeconomic status and adult earnings. These findings demonstrate that childhood waterborne lead exposure imposes substantial, persistent economic costs, with implications for cost-benefit analyses of environmental contamination and ongoing infrastructure remediation.

Lead Exposure | Labor Markets | Occupational Sorting

Introduction

Lead exposure is one of the most significant environmental health threats in American history. Millions of children were exposed through municipal water systems, leaded gasoline, lead-based paint, and contaminated soil throughout the twentieth century (1). Policy efforts to reduce exposure, including the phaseout of leaded gasoline, lead paint bans, and lead service line replacement, have generated a large literature documenting consequences of childhood lead exposure for cognition, behavior, health, birth outcomes, juvenile delinquency, educational attainment, and mortality (2–12). However, the economic burden of childhood exposure carried by affected cohorts across their adult working years, and the margins along which it falls, have not been fully established. How childhood lead exposure reshapes workers' careers across multiple dimensions of the labor market, whether the damage extends beyond what educational attainment captures, and whether childhood family resources can offset the harm are questions that require individual-level data linking adult wages, employment, and detailed occupations with childhood lead exposure and family socioeconomic status.

This paper uses a historical setting where exposed cohorts can be observed well into their working lives. During the late nineteenth and early twentieth centuries, American towns varied widely in the pipe materials used for municipal water delivery (13). Historical infrastructure records classify towns as using lead, a mixture of lead and other materials, or non-lead piping based on the materials reported in use at the time. Towns that adopted lead pipes tended to be larger and more industrial, a pattern of selection on observables documented in prior work (11, 13–15). The amount of lead leached from pipes into drinking water depended on local water chemistry, particularly pH and mineral content, which are determined by regional geology and stable over time (15–17). This interaction between pipe material and water chemistry creates a natural experiment: even among towns with the same pipe infrastructure, childhood lead exposure varied based on geologically determined water chemistry rather than municipal decisions or household choices. We link approximately 6.2 million men observed in the 1940 Census (ages 20–50) back to their

Significance Statement

Millions of American children were exposed to lead through municipal water pipes in the late nineteenth and early twentieth centuries, and aging lead infrastructure remains in service today. While lead's effects on health, cognition, and mortality are well established, its broader labor market consequences are less understood. We link 6.2 million men from the 1940 Census to their childhood towns and exploit geologically determined water chemistry that affects lead leaching only in towns with lead pipes. Men exposed as children earned lower wages, faced higher unemployment, and worked in lower-paying occupations decades into adulthood. Higher parental socioeconomic status did not protect them. Accounting for these labor market costs reveals larger returns to lead remediation than health effects alone capture.

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childhood town of residence at ages 0–10, when developing nervous systems are most vulnerable to lead’s effects (18–22). Combining these linked records with newly digitized data on historical pipe materials and water chemistry, we extend this established empirical framework from its prior focus on infant mortality and cognition to labor market outcomes measured at the individual level.

Comparing across pipe types, men who grew up in lead pipe towns with corrosive water earned 4.7% lower wages as adults, worked 0.67 fewer weeks per year, and were 1.1 percentage points more likely to be unemployed than men who grew up in non-lead pipe towns with similarly corrosive water. The earnings penalty reflects both where men work and how much they earn conditional on occupation: men from lead pipe towns were 2 percentage points more likely to work in below-median-wage occupations, and even within the same detailed occupation they earned 2.6% less than men from non-lead pipe towns, suggesting that lead exposure affected valued worker attributes beyond its effect on occupational sorting. Men from lead pipe towns show lower rates of childhood school attendance but no differences in self-reported years of completed education as adults.

The more tightly identified comparison exploits chemistry variation within pipe type. Among men who grew up in lead pipe towns, those in high leaching risk water environments earned 3.6% less than those in low leaching risk environments, while men in non-lead pipe towns show no differential outcomes across the same chemistry variation. This within-pipe-type contrast also reveals that corrosive water chemistry compresses the relationship between childhood socioeconomic status and adult earnings, reducing earnings for sons of middle- and upper-quartile fathers while having little effect on sons from the lowest quartile. Higher parental socioeconomic status did not shield children from the labor market harm caused by lead exposure.

This study contributes to the literature on lead exposure and economic outcomes in three ways. First, individual-level microdata allow us to examine the earnings penalty along two margins: occupational sorting and within-occupation wage losses, an analysis not possible with the town-level comparisons used in most historical studies of waterborne lead (13, 14, 23). By looking at multiple dimensions of individuals’ labor market outcomes, these findings provide a measure of the costs exposed cohorts carry into adulthood. Second, our historical setting allows for a test of whether higher parental socioeconomic status can buffer children from environmental harm. Because the health risks of lead pipes were unknown during this period (24), families within a given town did not selectively avoid exposure, so heterogeneity by family background among households with comparable connection to municipal water largely reflects differences in the capacity to buffer against a common exposure rather than differences in exposure itself. The within-pipe-type chemistry contrast, which compares otherwise similar towns that differ only in geologically determined water chemistry, provides the tightest test of this buffering question. Finally, we contribute to the conversation on the impacts of lead exposure on educational attainment. We find that the earnings penalty is not accompanied by differences in self-reported years of completed education when individuals are adults, but is accompanied by a reduced likelihood of being in school as

a child. Self-reported schooling in Census data is subject to heaping at grade-level milestones, and the institutional context of the High School Movement likely pushed children through grades regardless of cognitive capacity (25, 26). In our context, the earnings penalty appears to operate through channels that self-reported educational attainment may not capture, highlighting the value of directly observing labor market outcomes where possible.

Results

Geographic Variation in Waterborne Lead Exposure.

As American cities expanded during the late nineteenth century, lead emerged as a preferred material for water service pipes due to its durability and malleability, with lead pipes typically lasting twice as long as iron or steel alternatives (13). Nineteenth-century engineers and health officials actively encouraged lead plumbing, lacking understanding of low-level lead exposure risks (24, 27, 28). Engineers argued that protective mineral coatings would quickly form within pipes and eliminate leaching risks (29, 30), though such coatings actually required decades to develop and provided incomplete protection (15, 24). The result was wide variation in pipe adoption across towns (Figure 1): of the 971 towns in our sample, 229 (23%) used lead pipes, 281 (29%) used a mixture of lead and other materials, and 461 (48%) used no lead piping. Together, these towns contained municipal water systems serving roughly 35% of the US population in 1900. After conditioning on categorical fixed effects for region, population, industry composition, metropolitan status, and other town-level attributes, towns with different pipe types are similar on a wide range of 1900 Census characteristics (SI Appendix, Table 3).

Water chemistry determines the extent to which lead leaches from pipes into drinking water, a property known as “plumbosolvency” (16, 31, 32). Low pH combined with low mineral content produces corrosive conditions that substantially increase lead dissolution. In less corrosive environments, the protective mineral scales that engineers had predicted do eventually form on pipe interiors, reducing but not eliminating lead leaching. We classify each town as High or Low Leaching Risk based on historical measurements of pH and water hardness, and cross this with the three pipe material types to yield six exposure cells (SI Appendix, Table 1). The highest-exposure group, Pure Lead × High Leaching Risk, contains 55 towns. All estimates compare outcomes across these six cells, with Non-Lead × High Leaching Risk as the omitted category. This structure yields two types of comparisons: across pipe types within the same chemistry environment (e.g., Pure Lead vs. Non-Lead towns, both with High Leaching Risk water), and within pipe type across chemistry environments (e.g., Pure Lead towns with High vs. Low Leaching Risk water). The second comparison is more tightly identified because it holds town-type characteristics fixed and isolates the effect of geologically determined water chemistry on lead dissolution. Complete details on data sources, variable construction, and the retrospective linking methodology are provided in the SI Appendix.

Employment and Earnings Impacts. Figure 2 presents estimates of the effect of childhood waterborne lead exposure on adult labor market outcomes. Each panel plots coefficients for the five treatment cells relative to the

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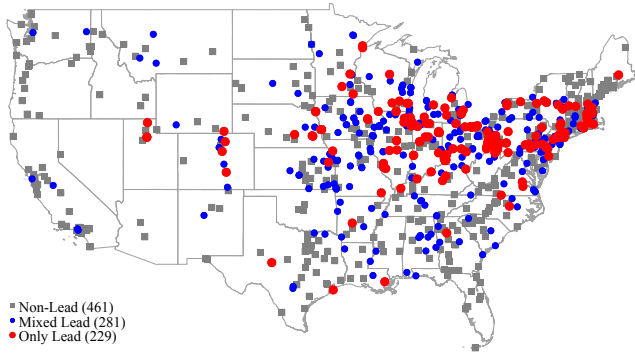


Fig. 1. Geographic distribution of the 971 towns in our analytical sample by water pipe material type. Towns are classified based on infrastructure records from *The Manual of American Water-Works* (1889–1897).

omitted category, Non-Lead towns in High Leaching Risk water environments (equation 1). All specifications absorb categorical fixed effects for the town-level attributes described in SI Appendix Section 3, as well as individual characteristics including age, race, household head's sex, occupation, and marital status, parental immigration status, homeownership, urban status, and group quarters status. Inverse probability weights account for differential selection into the linked sample, and standard errors are clustered at the pipe district level. Full regression results and variable definitions are provided in SI Appendix, Table 5 and Section 3.

Panel (A) of Figure 2 shows that men who grew up in Pure Lead \times High Leaching Risk towns earned 4.7% lower wage and salary income than men from Non-Lead \times High Leaching Risk towns ($p < 0.01$), conditional on positive earnings in the prior year. Panel (B) shows this same group of men worked 0.67 fewer weeks in the prior year ($p < 0.01$). Panel (C) shows they were 1.1 percentage points more likely to be unemployed at the time of enumeration ($p < 0.01$), a 13.4% increase relative to the 8.2% unemployment rate in Non-Lead towns. We find no effect on labor force participation.

The within-pipe-type comparison provides a more tightly identified estimate. Among men who grew up in Pure Lead towns, those in High Leaching Risk water environments earned 3.6% less than those in Low Leaching Risk environments ($p < 0.05$), worked nearly a full week less per year ($p < 0.01$), and were 1.1 percentage points more likely to be unemployed ($p < 0.01$) (SI Appendix, Table 5, Panel B). The same chemistry variation produces no differential effect on any outcome in Non-Lead towns, confirming that water chemistry affects labor market outcomes only in the presence of lead infrastructure. We do not find statistically significant effects for Mixed Lead towns in either comparison, consistent with lower average exposure from partial lead pipe coverage.

Occupational Sorting and Within-Occupation Wages. To better understand the earnings penalty, we examine two margins along which lead exposure could affect labor market outcomes: the types of occupations men hold and their earnings within those occupations. We classify occupations into two tiers based on occupation-level median earnings in 1940 and estimate whether lead exposure shifts the probability of working in a below-median occupation. Panel (A) of Figure 3 shows that men from Pure Lead \times High Leaching Risk towns were 2.0 percentage points more

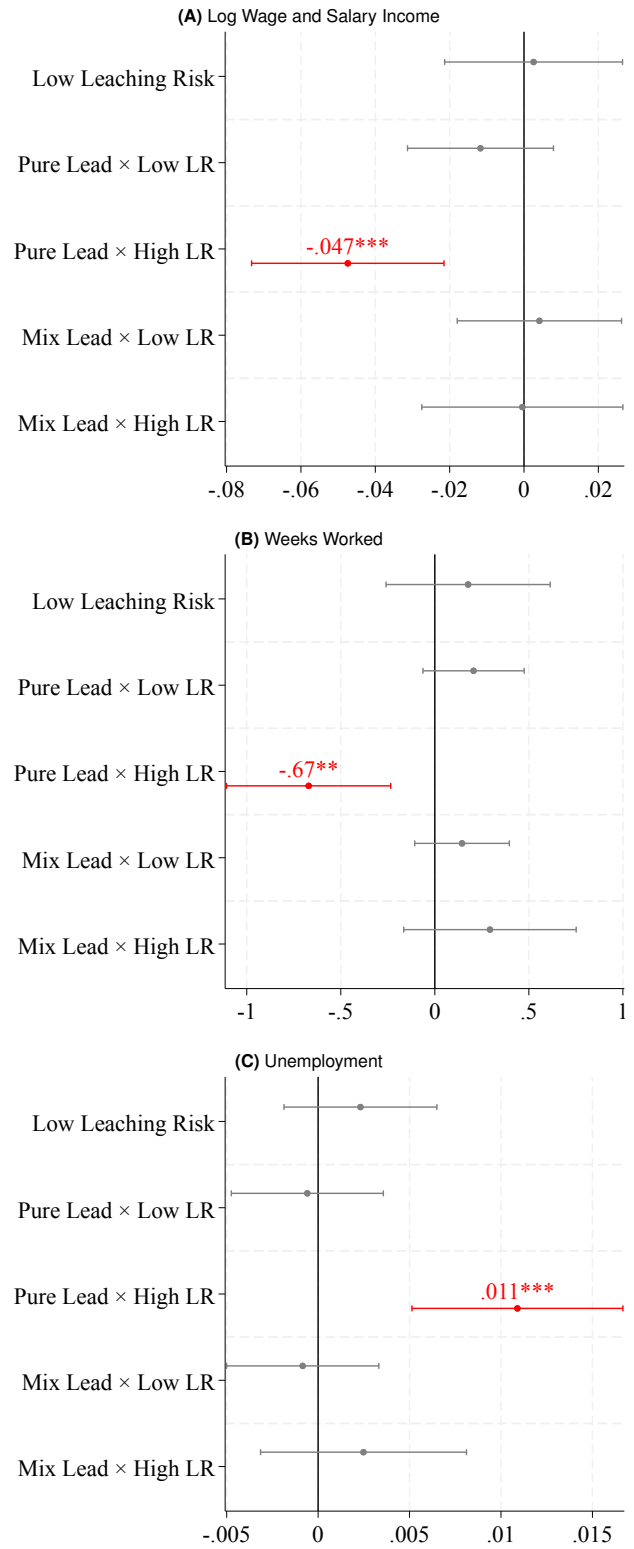


Fig. 2. Impact of childhood waterborne lead exposure on adult labor market outcomes. Each panel shows coefficient estimates and 90% confidence intervals from the primary IPW-weighted specification. The omitted category is Non-Lead pipes in High Leaching Risk environments. Pure Lead \times High LR indicates towns with lead pipes and high lead leaching risk water chemistry; Pure Lead \times Low LR indicates lead pipe towns with low leaching risk environments. Mixed Lead categories indicate towns using a combination of lead and other pipe materials. Standard errors clustered at the pipe district level. * $p < 0.10$, ** $p < 0.05$, *** $p < 0.01$.

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likely to work in lower-tier occupations than men from Non-Lead \times High Leaching Risk towns ($p < 0.01$). Consistent with the earnings results, the within-pipe-type chemistry contrast is significant for Pure Lead towns and near zero for Non-Lead towns, and we find no significant effects of Mixed Lead pipe exposure on occupational sorting.

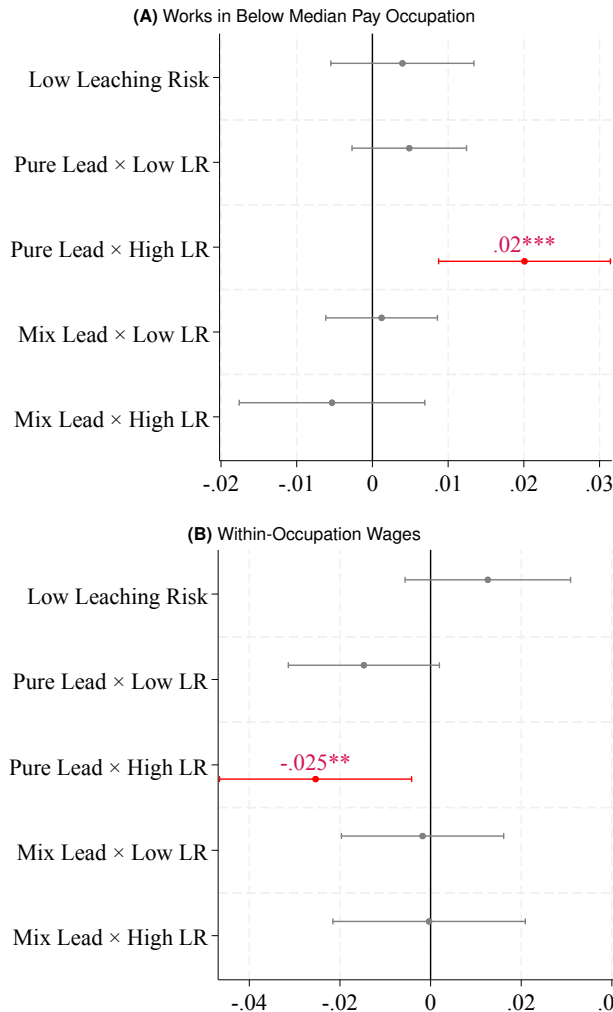


Fig. 3. Impact of childhood waterborne lead exposure on occupational sorting and within-occupation wages. In Panel (A), the dependent variable is an indicator for working in a lower-tier occupation, defined as below the occupation-level median earnings in 1940. Panel (B) augments equation (1) with 223 detailed occupation code fixed effects. Both panels show coefficient estimates and 90% confidence intervals from the primary IPW-weighted specification. The omitted category is Non-Lead pipes in High Leaching Risk water environments. Standard errors clustered at the pipe district level. * $p < 0.10$, ** $p < 0.05$, *** $p < 0.01$.

A key advantage of our individual-level microdata over town-level aggregates is the ability to examine wage differences within occupation categories. Panel (B) augments equation (1) with fixed effects for 223 detailed 1950-basis occupation codes, comparing earnings among men in the same occupation. Men from Pure Lead \times High Leaching Risk towns earned 2.6% less than men from Non-Lead \times High Leaching Risk towns working in the same detailed occupation ($p < 0.05$). The within-occupation penalty is larger in High Leaching Risk environments than Low Leaching Risk environments, and the Non-Lead chemistry contrast remains

near zero, confirming that the penalty operates through the chemistry channel. Because occupational placement is itself affected by lead exposure, the within-occupation estimate should be interpreted as descriptive: it indicates that the earnings penalty is not fully accounted for by the types of jobs men hold, but does not isolate a distinct causal channel. The finding is consistent with evidence that childhood lead exposure affects workplace-relevant skills and behaviors (15, 33, 34), though we cannot directly observe these pathways in the historical Census data.

Education. Despite the well-documented effects of lead on cognition and academic performance, we find no statistically significant effect of waterborne lead exposure on self-reported years of completed schooling in 1940 (SI Appendix, Figure 10, Panel (B)). We do find that children in Pure Lead \times High Leaching Risk towns were less likely to be attending school at ages 6–10 ($p = 0.065$) (SI Appendix, Figure 10, Panel (A)). Self-reported schooling in Census data is subject to heaping at grade-level milestones, which may obscure real differences in attainment, and the null result does not rule out differences in the quality of learning during the same reported years of schooling.

Can Household Resources Buffer Against Lead Exposure? A central finding in the contemporary lead literature is that disadvantaged children bear a disproportionate burden of lead exposure. Today, African American and low-income children are two to three times as likely to have elevated blood lead levels (6), lead service lines are disproportionately located in lower-income communities (35, 36), and the long-run effects of childhood lead exposure on adult outcomes are larger for children from low-income households (8). These disparities reflect two interacting forces: disadvantaged families face greater exposure, and they may have fewer resources to buffer the developmental consequences of whatever exposure occurs. In modern settings, these forces are difficult to separate because exposure and household resources are strongly correlated.

Our historical setting offers an opportunity to isolate the buffering question. Lead service pipes were concentrated in large and relatively wealthy American cities (13, 14, 23), and because the health risks were not understood (24), entire city populations consumed water delivered through the same infrastructure with no avoidance behavior. Unlike the contemporary pattern, where wealthier families can reduce exposure through residential choices and home remediation (6), selective avoidance in our setting is unlikely. Heterogeneity we observe by household socioeconomic status among households with comparable connection to municipal water therefore largely reflects differences in the capacity to buffer against a common exposure rather than differences in exposure itself.

We examine heterogeneity by childhood socioeconomic status, proxied by father's predicted income quartile rank when the individual was ages 0–10 (measurement details in SI Appendix, Section 5.4). Race and immigrant background are natural additional dimensions suggested by the contemporary literature. However, nonwhite men are nearly absent from lead-pipe cities in our linked sample, with fewer than 3% of men in Pure Lead \times High Leaching Risk towns being nonwhite, leaving insufficient variation to estimate differential effects by race (SI Appendix, Section 6.2). We find no differential effect by immigrant parentage, consistent with

497 the historical exposure pattern in which all residents of a
 498 lead-pipe city faced similar waterborne lead regardless of
 499 national origin.

500 Figure 4 displays earnings by father's income quartile
 501 for Pure Lead and Non-Lead towns, estimated using across-
 502 town variation with the same town-level controls as the main
 503 specification. Mixed Lead results are reported in the SI
 504 Appendix. All coefficients are measured relative to a single
 505 baseline group (Non-Lead \times High Leaching Risk \times Q1), so
 506 both levels and gradients are directly comparable across town
 507 types (SI Appendix, Section 5.4). Because lead-pipe cities
 508 were wealthier and more industrial, the composition of father's
 509 income quartiles differs across town types: Q1 fathers account
 510 for 17% of the Non-Lead \times High Leaching Risk population but
 511 only 5% of the Pure Lead \times High Leaching Risk population.
 512 Comparing gradients across town types in Panels (A) and
 513 (B) must be interpreted with this compositional difference
 514 in mind. Panel (C) avoids this concern by comparing within
 515 Pure Lead towns across chemistry environments, where the
 516 Q1 share is similar (5–6% in both leaching risk groups).

517 Panel (A) shows that in High Leaching Risk environments,
 518 the two pipe types start at comparable Q1 earnings levels,
 519 confirming that the baseline is similar after conditioning on
 520 town characteristics. The Non-Lead gradient rises steeply: Q4
 521 sons earn 10% more than Q1 sons ($p < 0.01$). The Pure Lead
 522 gradient is compressed: Q2 and Q3 sons show no earnings
 523 advantage over Q1, and Q4 sons earn only 4% more ($p < 0.05$).
 524 Panel (B) shows that in Low Leaching Risk environments,
 525 both pipe types display upward-sloping gradients of similar
 526 magnitude, confirming that lead infrastructure alone does
 527 not compress the SES gradient. The compression is specific
 528 to environments where chemistry promotes lead dissolution.

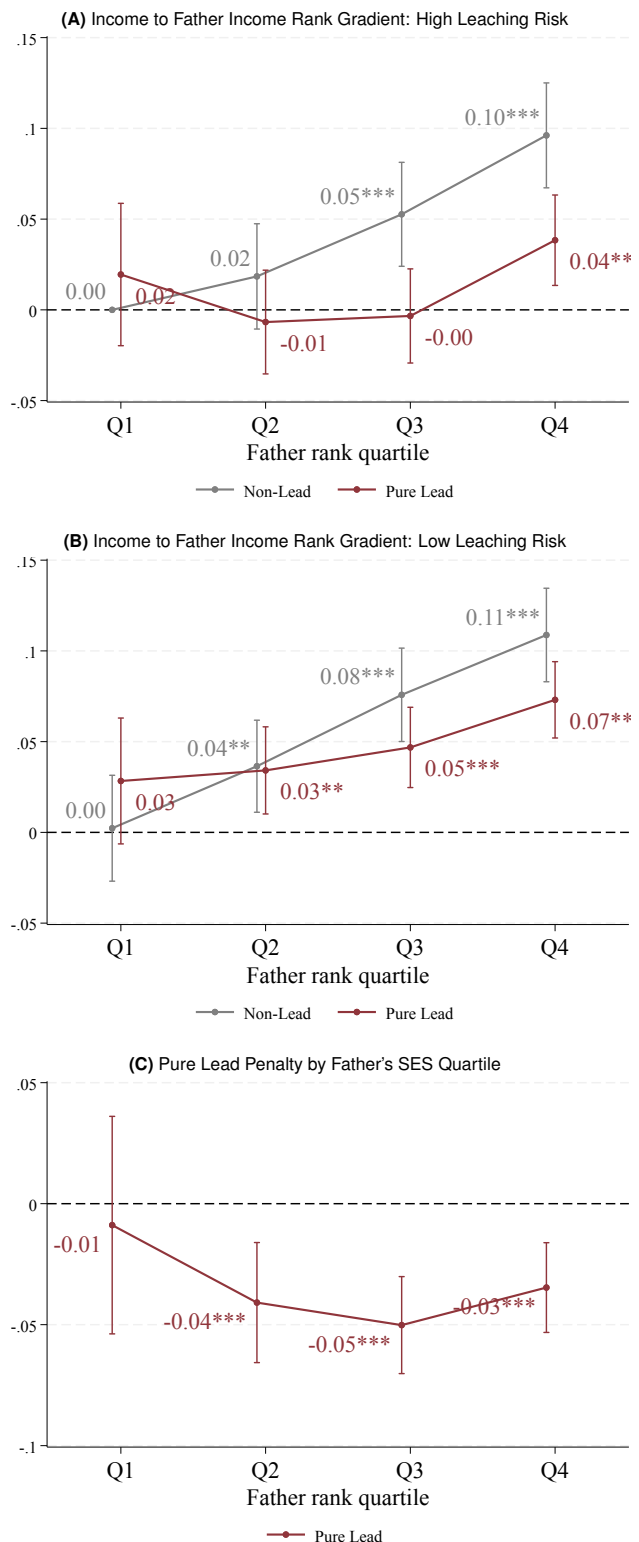
529 Panel (C) isolates the chemistry effect within Pure Lead
 530 towns by comparing High versus Low Leaching Risk
 531 environments at each quartile. Corrosive water chemistry
 532 reduces earnings at Q2 (–4%, $p < 0.01$), Q3 (–5%, $p < 0.01$),
 533 and Q4 (–3%, $p < 0.01$), with a small and insignificant
 534 effect at Q1. The Non-Lead chemistry contrast is near zero
 535 at every quartile (SI Appendix, Figure 9), confirming that
 536 water chemistry affects earnings only in the presence of lead
 537 pipes. Higher parental socioeconomic status did not shield
 538 children from the labor market harm caused by lead exposure
 539 in corrosive water environments: Q2, Q3, and Q4 sons all
 540 face significant earnings penalties while Q1 sons do not.

542 Discussion

543 Our findings establish that childhood waterborne lead expo-
 544 sure caused lasting labor market penalties. Men from lead
 545 pipe towns with corrosive water earned less, worked fewer
 546 weeks, and faced higher unemployment, with the within-pipe-
 547 type chemistry contrast confirming that these effects are
 548 driven by lead leaching rather than other correlates of urban
 549 infrastructure. The penalty operates along multiple margins
 550 of the labor market and is not offset by higher parental
 551 socioeconomic status.

553 Mechanisms Underlying Labor Market Penalties.

554 The labor market penalties we document are not accompanied
 555 by differences in self-reported years of completed education,
 556 despite evidence from this same historical period that lead
 557 exposure reduced cognitive test scores (15). The birth
 558 cohorts in our sample (1890–1920) attended school during



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Fig. 4. Earnings by father's income quartile and pipe type. All coefficients are relative to Non-Lead \times High LR \times Q1, estimated using across-town variation with town-level controls and IPW weights. Panel (A): High Leaching Risk environments. Panel (B): Low Leaching Risk environments. Panel (C): Within-pipe-type chemistry contrast for Pure Lead towns (High LR minus Low LR at each quartile); Non-Lead placebo near zero at all quartiles (see SI Appendix). 90% confidence intervals. Standard errors clustered at the pipe district level. * $p < 0.10$, ** $p < 0.05$, *** $p < 0.01$.

621 the High School Movement, a period in which secondary
622 schooling expanded rapidly, driven by local institutional
623 investment, compulsory attendance laws, and community
624 wealth (25, 26). In this institutional environment, modest
625 cognitive disadvantages attributable to lead exposure may
626 not have translated into differences in reported attainment,
627 particularly given that self-reported schooling in Census data
628 is subject to heaping at grade-level milestones. The finding
629 that lead-exposed men earn less even within the same detailed
630 occupation is consistent with a broader literature linking
631 childhood lead exposure to reduced cognitive functioning
632 (11, 15, 20, 37), as well as behavioral and conduct problems
633 that persist into adulthood (3, 33, 34). These capacities shape
634 workplace performance and career progression in ways that
635 Census data cannot directly observe, but that are consistent
636 with the pattern of results we find.

637 **Magnitude of Effects and Comparison to Prior**
638 **Research.** Our estimates are conservative for several
639 reasons. First, our analysis captures only waterborne lead
640 exposure. Men in our sample may also have been exposed
641 through lead-based paint and occupational contact, meaning
642 total lead exposure likely exceeded what our infrastructure-
643 based measure reflects. Second, our intent-to-treat design
644 assigns exposure based on town-level infrastructure and water
645 chemistry rather than individual blood lead levels. Actual
646 exposure varied within towns due to differences in water
647 consumption, duration of residence, and household plumbing,
648 so our estimates average across the exposure spectrum within
649 treated towns. Third, any remaining endogeneity in pipe-
650 type adoption is likely to bias our estimates toward zero.
651 Because larger and more economically developed cities were
652 more likely to adopt lead pipes, children raised in those
653 cities would plausibly have experienced better labor market
654 outcomes even in the absence of lead exposure, so simple
655 comparisons across pipe types tend to attenuate the estimated
656 negative effects of lead.

657 Our paper complements two recent papers exploring the
658 relationship between earnings and lead exposure using modern
659 policy variation. Reductions in airborne lead exposure for
660 in-utero children from the U.S. Clean Air Act increased adult
661 earnings by 3.5% (9). In Sweden, the phaseout of leaded
662 gasoline increased earnings by an estimated 4%, though
663 this is imputed from changes in educational attainment
664 rather than directly observed wages (8). Our estimates
665 are broadly comparable in magnitude to these other studies
666 despite measuring a different exposure source (waterborne vs.
667 airborne) in a different context, and despite the conservative
668 features of our intent-to-treat design.

669 **Heterogeneous Effects and the Compression of**
670 **Socioeconomic Advantages.** The compression of the
671 intergenerational earnings gradient in lead-exposed towns
672 speaks to a broader question about whether family resources
673 can protect children from environmental harm. Our cleanest
674 evidence is the within-pipe-type chemistry contrast (Figure 4,
675 Panel C), which holds pipe material fixed and compares
676 towns that differ only in geologically determined water
677 chemistry: corrosive chemistry in Pure Lead towns reduces
678 earnings for Q2, Q3, and Q4 sons, while the same chemistry
679 variation has no effect in Non-Lead towns. Across-pipe-type
680 comparisons at each quartile (Panels A and B) are broadly
681 consistent but rely on our covariates fully absorbing cross-

683 town compositional differences, which residual within-quartile
684 variation in father's occupation mix suggests they do not fully
685 do. One interpretation of the pattern is that lead exposure
686 reduces the returns to the human capital investments that
687 higher-SES families typically make in their children (38, 39):
688 if lead impairs the cognitive and behavioral capacities needed
689 to benefit from better schools, stronger networks, and greater
690 parental investment, then the advantages these resources
691 confer are diminished. We interpret the Q1 null with
692 caution: both differential mortality selection prior to the 1940
693 observation (12) and unobserved within-town differences in
694 actual lead exposure could contribute. Neither channel is
695 directly testable in our data. The broader pattern of results —
696 lead reduces earnings without reducing reported educational
697 attainment — is consistent with the interpretation that the
698 harm operates through capacities that family resources cannot
699 easily replace.

700 **Limitations.** Several limitations warrant consideration.
701 First, we cannot directly observe individual cognitive ability,
702 behavioral outcomes, or other valued worker attributes in
703 the Census data. The within-occupation wage penalty is
704 consistent with lead-induced reductions in these capacities,
705 but we cannot distinguish this from other sources of within-
706 occupation earnings variation such as firm quality, work
707 intensity, or chronic health conditions caused by lead exposure
708 itself. Second, we observe each individual in a town at
709 ages 0–10 but do not know the duration or precise timing
710 of their exposure. The medical literature indicates that
711 in-utero and early childhood exposure produce the largest
712 developmental effects, but we cannot isolate these critical
713 windows from later childhood exposure. Third, our analysis
714 focuses exclusively on men because women's labor market
715 participation was limited in this period. The effects of lead
716 exposure on women likely operate through other margins,
717 including fertility, marriage, and household production, that
718 we do not examine. Fourth, nonwhite men are nearly absent
719 from lead-pipe cities in our linked sample, preventing us
720 from examining racial heterogeneity in the effects of lead
721 exposure. Given the prominent role of racial disparities in the
722 modern lead literature (6), this remains an important question
723 that our data cannot address. Fifth, while conditional
724 differences in household head characteristics across pipe types
725 are small after town-level controls (SI Appendix, Table 4),
726 parents who chose to live in lead-pipe cities may have differed
727 on dimensions we cannot observe. The within-pipe-type
728 chemistry contrast mitigates this concern because it compares
729 families within the same type of city, and water chemistry
730 was geologically determined and not a factor in household
731 location decisions. Sixth, our inverse probability weights
732 account for mortality between an individual's first census
733 observation and 1940, but the direction of any remaining
734 mortality selection is ambiguous: if the most severely affected
735 individuals died before observation, our estimates understate
736 the true effect, but surviving individuals may also carry
737 chronic health burdens that contribute to the labor market
738 penalties we measure.

739 **Implications for Policy and Future Research.** These
740 findings are relevant for ongoing efforts to evaluate the
741 costs and benefits of lead remediation. Current cost-benefit
742 analyses of lead service line replacement focus primarily on
743 health outcomes and cognitive gains. Our results indicate
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745 that the economic costs of lead exposure extend to labor
746 market outcomes measured decades after childhood exposure,
747 including reduced earnings, higher unemployment, and
748 occupational downgrading. Accounting for these labor market
749 costs alongside health effects provides a more complete picture
750 of the returns to lead remediation, consistent with other
751 work documenting broad economic returns to investments in
752 water and sanitary infrastructure (40). We estimate annual
753 per-person earnings losses of \$950 to \$1,350 in 2026 dollars,
754 combining the wage penalty and increased unemployment
755 (see SI Appendix, Section 6.3 for details). Over the 30-
756 year working life observed in our data, cumulative losses
757 range from \$28,500 to \$40,500 per exposed individual. The
758 EPA estimates a central cost of \$4,700 per full lead service
759 line replacement, with a range of \$1,200 to \$12,300 (36).
760 The per-person earnings losses we estimate exceed even the
761 high end of replacement costs by a factor of two to three,
762 and these losses reflect only male earnings from a single
763 exposure source, omitting health costs, mortality, and effects
764 on women. The SES findings add a distributional dimension:
765 lead exposure compresses the returns to family background,
766 suggesting that cost-benefit frameworks focusing only on
767 average treatment effects may miss important consequences
768 for economic mobility. Future research should prioritize
769 examining the effects of lead exposure on women, on racial
770 and ethnic minorities where data permit, and on outcomes
771 later in the life course including health, disability, and
772 retirement security (1).

774 Materials and Methods

775 **Data Sources.** The empirical analysis combines three primary
776 data sources. First, individual-level data from the 1940 U.S.
777 Decennial Census provide information on adult labor market
778 outcomes (41). Second, we use the full-count 1900, 1910, and
779 1920 U.S. Decennial Censuses with person identifiers (HistIDs)
780 provided by IPUMS (41) and cross-decade linkages from the
781 Census Tree (42, 43). Third, hand-digitized historical sources
782 on municipal water infrastructure and chemistry come from *The*
783 *Manual of American Water-Works*, which documents service pipe
784 materials at the town level, and U.S. Geological Survey publications
785 documenting historical measurements of pH and hardness (44, 45).
786 Full details on data sources, digitization procedures, and linkage
787 methodology and rates are provided in SI Appendix, Section 1.

788 **Sample Construction.** Our sample is males ages 20–50 in the
789 1940 Census who can be linked to their childhood census record
790 and who can be linked to town-level water infrastructure and
791 chemistry information. We restrict the analysis to men because
792 women’s labor market participation was limited in this period. We
793 link each individual back to their childhood Census record when
794 they were ages 0–10, so that each individual in the sample has
795 an initial observation (in either 1900, 1910, or 1920) and a 1940
796 observation. These sampling criteria yield 6.2 million men from
797 971 unique towns spanning all regions of the United States.

798 **Measuring Childhood Lead Exposure.** We categorize
799 towns into three pipe material groups based on infrastructure
800 records: Pure Lead (lead reported as the pipe material in use),
801 Mixed Lead (a combination of lead and other materials reported),
802 and Non-Lead (no lead pipes reported). To determine lead leaching
803 potential, we predict theoretical lead solubility using Schock’s
804 solubility curves, which describe lead concentration as a non-linear
805 function of pH and alkalinity (16, 31, 32). Because alkalinity is not
806 consistently available in historical water quality records, we follow
807 prior literature in using water hardness as a proxy and implement
808 a Generalized Additive Model (GAM) fitted to digitized data from
809 Schock’s solubility figure. Towns are classified as High or Low
810 Leaching Risk environments based on predicted lead solubility, with
811 the threshold set at the 75th percentile of predicted values across
812 all towns. This classification creates six exposure categories from

807 the cross-classification of pipe type and leaching environment. As
808 a robustness check, we provide the main results for two alternative
809 classifications of High and Low Leaching Risk environments, at
810 the 50th percentile of predicted lead values and at pH cutoffs
811 used in prior literature (11, 15), in SI Appendix Section 7. Full
812 technical details on the predicted lead solubility model, including
813 extrapolation strategies for extreme pH and hardness values, are
814 provided in SI Appendix, Section 2.

815 We provide evidence in SI Appendix, Section 4, that conditional
816 on the town-level fixed effects used in all specifications, remaining
817 differences in pre-existing town characteristics across pipe types
818 are small and statistically insignificant.

819 **Outcome Variables.** The primary outcome is wage and
820 salary income in 1939, constructed from 1940 Census questions
821 on annual earnings from wages and salaries (excluding self-
822 employment income, business profits, investment income, and
823 government transfers). We also examine weeks worked in the
824 previous year, unemployment status, labor force participation, and
825 occupational category. Occupations are classified using harmonized
826 1950 occupation codes provided by IPUMS. We create two broad
827 occupational tiers by splitting all occupations at the median of
828 average occupation-level earnings in 1940: lower-tier occupations
829 include laborers, service workers, and lower-skilled operatives, while
830 upper-tier occupations include professionals, managers, craftsmen,
831 and skilled operatives. All monetary values are expressed in 1940
832 dollars. Additional details on variable construction and treatment
833 of missing or top-coded values are provided in SI Appendix,
834 Section 3.

835 **Empirical Specification.** We estimate the effect of childhood
836 waterborne lead exposure on adult labor market outcomes using
837 ordinary least squares regressions of the form:

$$838 Y_{ic} = \beta_1 \text{LowLR}_c + \beta_2 \text{PureLead}_c \times \text{LowLR}_c$$

$$839 + \beta_3 \text{PureLead}_c \times \text{HighLR}_c + \beta_4 \text{MixLead}_c \times \text{LowLR}_c$$

$$840 + \beta_5 \text{MixLead}_c \times \text{HighLR}_c + X_i' \gamma + Z_c' \theta + \alpha_r + \varepsilon_{ic} \quad [1]$$

841 where Y_{ic} is the outcome for individual i who grew up in town
842 c . The omitted category is Non-Lead pipes in High Leaching Risk
843 (High LR) environments. The coefficient β_3 captures the effect of
844 Pure Lead pipes in High LR water, while β_1 measures the main
845 effect of Low LR environments for Non-Lead towns. Individual
846 controls (X_i) include age, race, household head’s sex, occupation,
847 and marital status, parental immigration status, homeownership,
848 urban status, and group quarters status. Region fixed effects (α_r)
849 control for time-invariant regional differences based on childhood
850 location. Town-level controls (Z_c) are absorbed as categorical fixed
851 effects for population, industry composition, metropolitan status,
852 proximity to large cities, water system characteristics, and mining
853 activity. Standard errors are clustered at the pipe district level,
854 which groups towns sharing a common water system (824 clusters).
855 Inverse probability weights are used to weight an individual back
856 to their representation in their original census observation (46).
857 Unweighted results are comparable and reported in SI Appendix 7.

858 For occupational sorting analysis, we estimate equation (1)
859 separately for an indicator of lower-tier occupation. For within-
860 occupation wage analysis, we augment equation (1) with 223
861 detailed occupation fixed effects, comparing earnings among
862 workers in the same occupation. For heterogeneity analysis by
863 childhood socioeconomic status, we interact equation (1) with
864 the four father income quartiles. Complete variable definitions,
865 regression specifications, and robustness checks are provided in SI
866 Appendix, Sections 3, 5, and 7.

867 **Validation of Research Design.** To support the validity
868 of our research design, we conduct several tests. First, we
869 show that conditional on the town-level fixed effects used in all
870 specifications, towns with different pipe materials are balanced on
871 pre-exposure demographic and economic characteristics observed
872 in the 1900 Census (SI Appendix, Table 3), and that household
873 head characteristics are similarly balanced at the individual level
874 (SI Appendix, Table 4). Second, we demonstrate that leaching
875 risk is not systematically correlated with pipe type choice (SI
876 Appendix, Figure 8). Third, we show that linkage rates from
877 childhood to adulthood do not vary systematically with pipe type
878

869 or leaching environment (SI Appendix, Figure 2). Finally, we
870 conduct robustness checks including alternative chemistry cutoffs,
871 state fixed effects, town-level leave-one-out estimates around the
872 highest exposure group, and excluding modern pipe assignments
873 (SI Appendix, Section 7).
874

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