

# Economic Consequences of Childhood Exposure to Urban Environmental Toxins

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**Childhood lead exposure causes well-documented damage to cognition, health, and mortality. Whether, and how, these harms translate into lasting labor market penalties is less clear. We estimate the consequences of childhood lead exposure on adult labor market outcomes by linking 6.2 million men, ages 20 - 50, from the 1940 full-count U.S. Census to their earliest childhood decennial census record across 971 towns. Identification comes from the interaction of municipal pipe materials with geologically determined water chemistry that causes corrosion, leading to elevated lead leaching. Comparing across pipe types, men who were ages 0 to 10 in pure lead pipe towns with corrosive water earned 4.7% lower wages, worked 0.67 fewer weeks per year, and were 1.1 percentage points more likely to be unemployed than men in non-lead towns. The more tightly identified comparison exploits chemistry variation within pipe type: among pure lead towns, men in high leaching risk water environments earned 3.6% less than those in low leaching risk environments, while the same chemistry variation has no differential effect on outcomes in towns without lead pipes. The wage penalty operates through both occupational sorting into lower-paying jobs and reduced earnings within the same detailed occupation. Lead exposure reduces childhood school attendance but does not affect self-reported completed years of education as of 1940. Elevated lead exposure also compresses the relationship between childhood socioeconomic status and adult earnings. These findings demonstrate that childhood waterborne lead exposure imposes substantial, lifelong economic costs through multiple pathways, which have implications for the full cost-benefit analysis of environmental contamination and the returns to ongoing infrastructure remediation.**

Lead Exposure | Labor Markets | Occupational Sorting

## Introduction

Lead exposure is one of the most significant environmental health threats in American history. Millions of children were exposed through municipal water systems, leaded gasoline, lead-based paint, and contaminated soil throughout the twentieth century (1). Policy efforts to reduce exposure, including the phaseout of leaded gasoline, lead paint bans, and lead service line replacement, have generated a large literature documenting consequences for cognition, behavior, health, juvenile delinquency, educational attainment, and mortality (2–7). However, we have largely overlooked the economic burden carried by generations exposed in the past. Beginning to understand the labor market consequences of historical lead exposure is critical for assessing the true societal costs of environmental contamination and informing ongoing lead remediation efforts.

This paper uses a historical setting where exposed cohorts can be observed well into their working lives. During the late nineteenth and early twentieth centuries, American towns varied widely in the pipe materials used for municipal water delivery (8). Historical infrastructure records classify towns as using lead, a mixture of lead and other materials, or non-lead piping based on the materials reported in use at the time. Towns that adopted lead pipes tended to be larger and more industrial, a pattern of selection on observables documented in prior work (8–11). The amount of lead leached from pipes into drinking water depended on local water chemistry, particularly pH and mineral content, which are determined by regional geology and stable over time (10, 12, 13). This interaction between pipe material and water chemistry creates a natural experiment: even among towns with the same pipe infrastructure, childhood lead exposure varied based on geologically determined water chemistry rather than municipal decisions or household choices. We link approximately 6.2 million men observed in the 1940 Census (ages 20–50) back to their childhood town of residence at ages 0–10, when developing nervous systems are most vulnerable to lead’s effects (14–18). Combining these linked records with newly digitized data on historical pipe materials and water chemistry, we extend this

## Significance Statement

Millions of American children were exposed to lead through municipal water pipes during the late nineteenth and early twentieth centuries, and aging lead infrastructure remains in service in cities across the United States. The health, cognitive, and mortality consequences of childhood lead exposure are well established, but less is known about how exposed children carry these burdens into the labor market as adults. We examine this question by linking 6.2 million men from the 1940 full-count Census to their childhood towns and exploiting geologically determined water chemistry variation that governs lead leaching from municipal pipes. This chemistry variation affects labor market outcomes only in towns with lead pipes, isolating lead exposure from other correlates of urban infrastructure. The resulting wage penalties, unemployment increases, and occupational downgrading persist decades after childhood exposure and are not offset by higher parental socioeconomic status. Accounting for these labor market costs alongside health effects provides a more complete picture of the returns to ongoing lead service line replacement.

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125 established empirical framework from its prior focus on infant  
126 mortality and cognition to labor market outcomes measured  
127 at the individual level.

128 Comparing across pipe types, men who grew up in lead  
129 pipe towns with corrosive water earned 4.7% lower wages  
130 as adults, worked 0.67 fewer weeks per year, and were 1.1  
131 percentage points more likely to be unemployed than men  
132 who grew up in non-lead pipe towns with similarly corrosive  
133 water. The earnings penalty reflects both where men work  
134 and how much they earn conditional on occupation: men from  
135 lead pipe towns were 2 percentage points more likely to work  
136 in below-median-wage occupations, and even within the same  
137 detailed occupation they earned 2.6% less than men from non-  
138 lead pipe towns, suggesting that lead exposure affected valued  
139 worker attributes beyond its effect on occupational sorting.  
140 Men from lead pipe towns show lower rates of childhood  
141 school attendance but no differences in self-reported years of  
142 completed education as adults.

143 The more tightly identified comparison exploits chemistry  
144 variation within pipe type. Among men who grew up in lead  
145 pipe towns, those in high leaching risk water environments  
146 earned 3.6% less than those in low leaching risk environments,  
147 while men in non-lead pipe towns show no differential  
148 outcomes across the same chemistry variation. This within-  
149 pipe-type contrast also reveals that corrosive water chemistry  
150 compresses the relationship between childhood socioeconomic  
151 status and adult earnings, reducing earnings for sons of  
152 middle- and upper-quartile fathers while having little effect on  
153 sons from the lowest quartile. Higher parental socioeconomic  
154 status did not shield children from the labor market harm  
155 caused by lead exposure.

156 This study contributes to the literature on lead exposure  
157 and economic outcomes in three ways. First, individual-level  
158 microdata allow us to examine the earnings penalty along  
159 two margins: occupational sorting and within-occupation  
160 wage losses, an analysis not possible with the town-level  
161 comparisons used in most historical studies of waterborne  
162 lead (8, 9, 19). (11) use a similar identification strategy  
163 to show that childhood waterborne lead exposure reduces  
164 cognitive functioning in older adults, but do not examine  
165 labor market outcomes. Second, we find that the earnings  
166 penalty is not accompanied by differences in self-reported  
167 years of completed education, despite evidence from this same  
168 historical setting that lead exposure reduced cognitive test  
169 scores (10). Self-reported schooling in Census data is subject  
170 to heaping at grade-level milestones, and the institutional  
171 context of the High School Movement likely pushed children  
172 through grades regardless of cognitive capacity (20, 21). In  
173 our context, the earnings penalty operates through channels  
174 that self-reported educational attainment does not capture,  
175 highlighting the value of directly observing labor market  
176 outcomes where possible. Third, our historical setting allows  
177 a clean test of whether higher parental socioeconomic status  
178 can buffer children from environmental harm. Because the  
179 health risks of lead pipes were unknown during this period  
180 (22), families could not selectively avoid exposure, so any  
181 heterogeneity by family background reflects differences in the  
182 capacity to buffer against a common exposure rather than  
183 differences in exposure itself. As policy efforts to replace  
184 remaining lead infrastructure continue, these findings provide  
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a measure of the labor market costs that exposed cohorts  
carry into adulthood.

## Results

### Geographic Variation in Waterborne Lead Exposure.

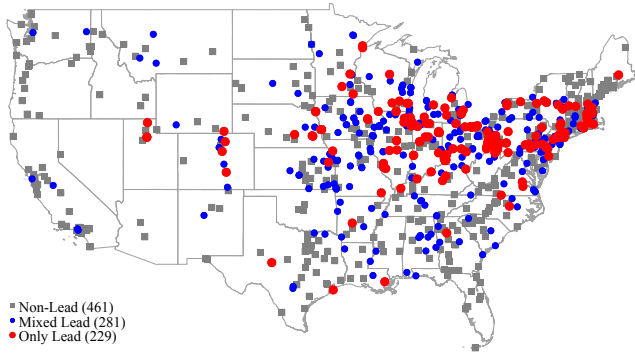
192 As American cities expanded during the late nineteenth  
193 century, lead emerged as a preferred material for water service  
194 pipes due to its durability and malleability, with lead pipes  
195 typically lasting twice as long as iron or steel alternatives  
196 (8). Nineteenth-century engineers and health officials actively  
197 encouraged lead plumbing, lacking understanding of low-  
198 level lead exposure risks (22–24). Engineers argued that  
199 protective mineral coatings would quickly form within pipes  
200 and eliminate leaching risks (25, 26), though such coatings  
201 actually required decades to develop and provided incomplete  
202 protection (10, 22). The result was wide variation in pipe  
203 adoption across towns (Figure 1): of the 971 towns in our  
204 sample, 229 (23%) used lead pipes, 281 (29%) used a mixture  
205 of lead and other materials, and 461 (48%) used no lead piping.  
206 After conditioning on categorical fixed effects for region,  
207 population, industry composition, metropolitan status, and  
208 other town-level attributes, towns with different pipe types  
209 are similar on a wide range of 1900 Census characteristics (SI  
210 Appendix, Table 3).

211 Water chemistry determines the extent to which lead  
212 leaches from pipes into drinking water, a property known as  
213 “plumbosolvency” (12, 27, 28). Low pH combined with low  
214 mineral content produces corrosive conditions that substan-  
215 tially increase lead dissolution. In less corrosive environments,  
216 the protective mineral scales that engineers had predicted do  
217 eventually form on pipe interiors, reducing but not eliminating  
218 lead leaching. We classify each town as High or Low Leaching  
219 Risk based on historical measurements of pH and water  
220 hardness, and cross this with the three pipe material types to  
221 yield six exposure cells (SI Appendix, Table 1). The highest-  
222 exposure group, Pure Lead × High Leaching Risk, contains 55  
223 towns. All estimates compare outcomes across these six cells,  
224 with Non-Lead × High Leaching Risk as the omitted category.  
225 This structure yields two types of comparisons: across pipe  
226 types within the same chemistry environment (e.g., Pure Lead  
227 vs. Non-Lead towns, both with High Leaching Risk water),  
228 and within pipe type across chemistry environments (e.g.,  
229 Pure Lead towns with High vs. Low Leaching Risk water).  
230 The second comparison is more tightly identified because it  
231 holds town-type characteristics fixed and isolates the effect of  
232 geologically determined water chemistry on lead dissolution.  
233 Complete details on data sources, variable construction, and  
234 the retrospective linking methodology are provided in the SI  
235 Appendix.

### Employment and Earnings Impacts.

237 Figure 2 presents estimates of the effect of childhood waterborne  
238 lead exposure on adult labor market outcomes. Each panel  
239 plots coefficients for the five treatment cells relative to the  
240 omitted category, Non-Lead towns in High Leaching Risk  
241 water environments (equation 1). All specifications absorb  
242 categorical fixed effects for the town-level attributes described  
243 above, as well as individual characteristics including age,  
244 race, household head’s sex, occupation, and marital status,  
245 parental immigration status, homeownership, and urban  
246 status. Inverse probability weights account for differential  
247 selection into the linked sample, and standard errors are  
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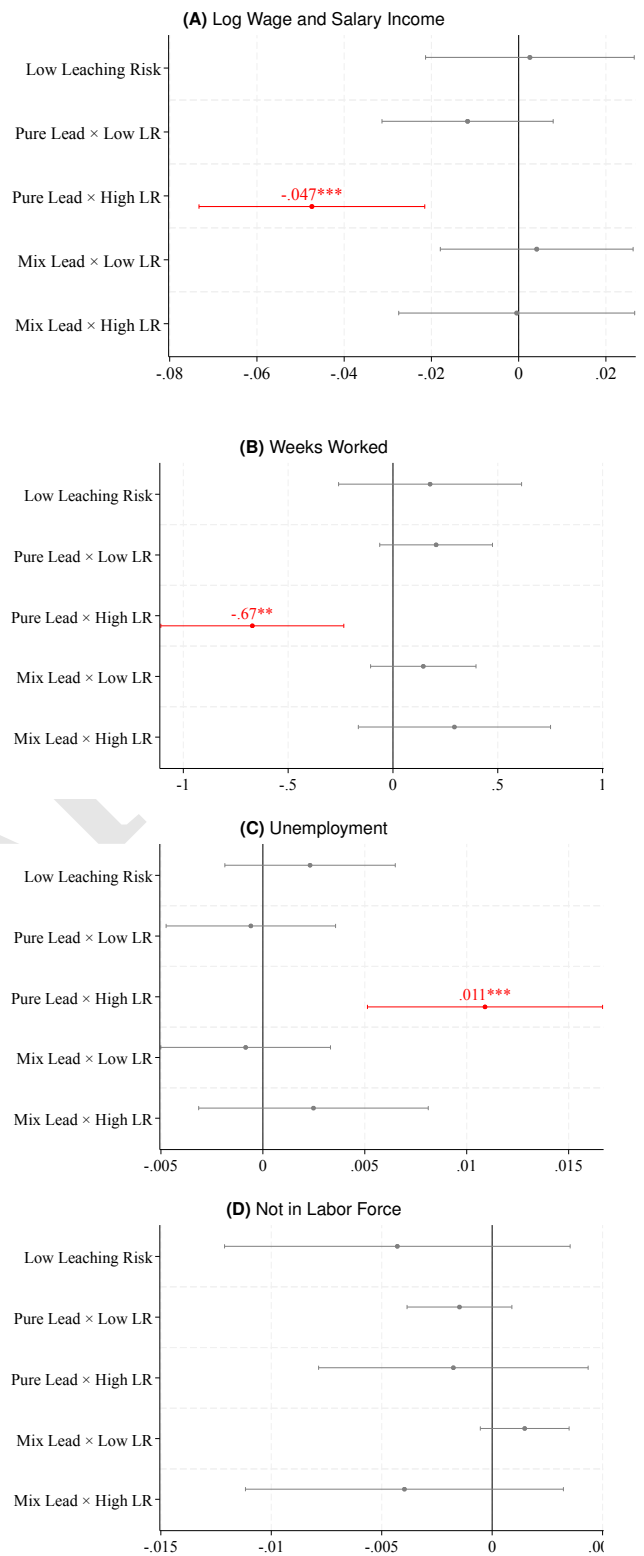
**Fig. 1.** Geographic distribution of the 971 towns in our analytical sample by water pipe material type. Towns are classified based on infrastructure records from *The Manual of American Water-Works* (1889–1897).

clustered at the pipe district level. Full regression results and variable definitions are provided in SI Appendix, Table 5 and Section 3.

Panel (A) shows that men who grew up in Pure Lead  $\times$  High Leaching Risk towns earned 4.7% lower wages than men from Non-Lead  $\times$  High Leaching Risk towns ( $p < 0.01$ ), where wages are log wage and salary income conditional on positive earnings in the prior year. Panel (B) shows these men worked 0.67 fewer weeks in the prior year, including zeros ( $p < 0.01$ ). Panel (C) shows they were 1.1 percentage points more likely to be unemployed at the time of enumeration ( $p < 0.01$ ), a 13.4% increase relative to the 8.2% unemployment rate in Non-Lead towns. Panel (D) shows no effect on labor force participation.

The within-pipe-type comparison provides a more tightly identified estimate. Among men who grew up in Pure Lead towns, those in High Leaching Risk water environments earned 3.6% less than those in Low Leaching Risk environments ( $p < 0.05$ ), worked nearly a full week less per year ( $p < 0.01$ ), and were 1.1 percentage points more likely to be unemployed ( $p < 0.01$ ) (SI Appendix, Table 5, Panel B). The same chemistry variation produces no differential effect on any outcome in Non-Lead towns, confirming that water chemistry affects labor market outcomes only in the presence of lead infrastructure. We do not find statistically significant effects for Mixed Lead towns in either comparison, consistent with lower average exposure from partial lead pipe coverage.

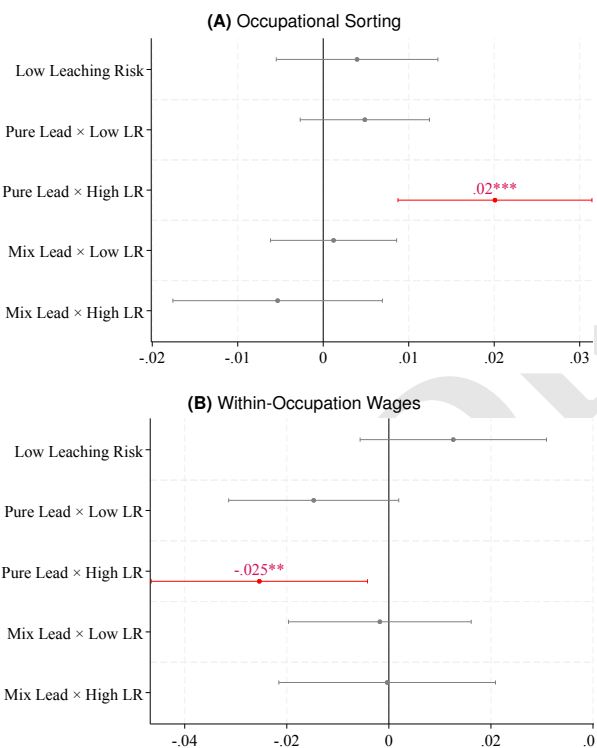
**Occupational Sorting and Within-Occupation Wages.** To better understand the earnings penalty, we examine two margins along which lead exposure could affect labor market outcomes: the types of occupations men hold and their earnings within those occupations. We classify occupations into two tiers based on occupation-level median earnings in 1940 and estimate whether lead exposure shifts the probability of working in a below-median occupation. Panel (A) of Figure 3 shows that men from Pure Lead  $\times$  High Leaching Risk towns were 2.0 percentage points more likely to work in lower-tier occupations than men from Non-Lead  $\times$  High Leaching Risk towns ( $p < 0.01$ ). Consistent with the earnings results, the within-pipe-type chemistry contrast is significant for Pure Lead towns and near zero for Non-Lead towns, and we find no significant effects of Mixed Lead pipe exposure on occupational sorting.



**Fig. 2.** Impact of childhood waterborne lead exposure on adult labor market outcomes. Each panel shows coefficient estimates and 90% confidence intervals from the primary IPW-weighted specification. The omitted category is Non-Lead pipes in High Leaching Risk environments. Pure Lead  $\times$  High LR indicates towns with lead pipes and high lead leaching risk water chemistry; Pure Lead  $\times$  Low LR indicates lead pipe towns with low leaching risk environments. Mixed Lead categories indicate towns using a combination of lead and other pipe materials. Standard errors clustered at the pipe district level. \* $p < 0.10$ , \*\* $p < 0.05$ , \*\*\* $p < 0.01$ .

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373 A key advantage of our individual-level microdata over  
 374 town-level aggregates is the ability to examine wage differ-  
 375 ences within occupation categories. Panel (B) augments  
 376 equation (1) with fixed effects for 223 detailed 1950-basis  
 377 occupation codes, comparing earnings among men in the  
 378 same occupation. Men from Pure Lead  $\times$  High Leaching Risk  
 379 towns earned 2.6% less than men from Non-Lead  $\times$  High  
 380 Leaching Risk towns working in the same detailed occupation  
 381 ( $p < 0.05$ ). The within-occupation penalty is larger in  
 382 High Leaching Risk environments than Low Leaching Risk  
 383 environments, and the Non-Lead chemistry contrast remains  
 384 near zero, confirming that the penalty operates through  
 385 the chemistry channel. Because occupational placement  
 386 is itself affected by lead exposure, the within-occupation  
 387 estimate should be interpreted as descriptive: it indicates  
 388 that the earnings penalty is not fully accounted for by the  
 389 types of jobs men hold, but does not isolate a distinct  
 390 causal channel. The finding is consistent with evidence that  
 391 childhood lead exposure affects workplace-relevant skills and  
 392 behaviors (29, 30), though we cannot directly observe these  
 393 pathways in the historical Census data.



421 **Fig. 3.** Impact of childhood waterborne lead exposure on occupational sorting and  
 422 within-occupation wages. In Panel (A), the dependent variable is an indicator for  
 423 working in a lower-tier occupation, defined as below the occupation-level median  
 424 earnings in 1940. Panel (B) augments equation (1) with 223 detailed occupation  
 425 code fixed effects. Both panels show coefficient estimates and 90% confidence  
 426 intervals from the primary IPW-weighted specification. The omitted category is Non-  
 427 Lead pipes in High Leaching Risk water environments. Standard errors clustered at  
 428 the pipe district level. \* $p < 0.10$ , \*\* $p < 0.05$ , \*\*\* $p < 0.01$ .

429 **Education.** Despite the well-documented effects of  
 430 lead on cognition and academic performance, we find no  
 431 statistically significant effect of waterborne lead exposure  
 432 on self-reported years of completed schooling in 1940 (SI  
 433 Appendix, Figure 10, Panel (B)). We do find that children

435 in Pure Lead  $\times$  High Leaching Risk towns were less likely to  
 436 be attending school at ages 6–10 ( $p = 0.065$ ) (SI Appendix,  
 437 Figure 10, Panel (A)). Self-reported schooling in Census data  
 438 is subject to heaping at grade-level milestones, which may  
 439 obscure real differences in attainment, and the null result  
 440 does not rule out differences in the quality of learning during  
 441 the same reported years of schooling.

442 **Can Household Resources Buffer Against Lead  
 Exposure?** A central finding in the contemporary lead  
 443 literature is that disadvantaged children bear a disproportio-  
 444 nate burden of lead exposure. Today, African American  
 445 and low-income children are two to three times as likely  
 446 to have elevated blood lead levels (5), lead service lines  
 447 are disproportionately located in lower-income communities  
 448 (31, 32), and the long-run effects of childhood lead exposure  
 449 on adult outcomes are larger for children from low-income  
 450 households (6). These disparities reflect two interacting forces:  
 451 disadvantaged families face greater exposure, and they may  
 452 have fewer resources to buffer the developmental consequences  
 453 of whatever exposure occurs. In modern settings, these forces  
 454 are difficult to separate because exposure and household  
 455 resources are strongly correlated.

456 Our historical setting offers an opportunity to isolate the  
 457 buffering question. Lead service pipes were concentrated  
 458 in large and relatively wealthy American cities (8, 9, 19),  
 459 and because the health risks were not understood (22),  
 460 entire city populations consumed water delivered through the  
 461 same infrastructure with no avoidance behavior. Unlike the  
 462 contemporary pattern, where wealthier families can reduce  
 463 exposure through residential choices and home remediation  
 464 (5), selective avoidance in our setting is unlikely. Any  
 465 heterogeneity we observe by household socioeconomic status  
 466 therefore reflects differences in the capacity to buffer against  
 467 a common exposure, not differences in exposure itself.

468 We examine heterogeneity by childhood socioeconomic  
 469 status, proxied by father’s predicted income quartile rank  
 470 when the individual was ages 0–10 (measurement details in SI  
 471 Appendix, Section 5.4). Race and immigrant background are  
 472 natural additional dimensions suggested by the contemporary  
 473 literature. However, nonwhite men are nearly absent from  
 474 lead-pipe cities in our linked sample, with fewer than 3%  
 475 of men in Pure Lead  $\times$  High Leaching Risk towns being  
 476 nonwhite, leaving insufficient variation to estimate differential  
 477 effects by race (SI Appendix, Section 6.2). We find no  
 478 differential effect by immigrant parentage, consistent with  
 479 the historical exposure pattern in which all residents of a  
 480 lead-pipe city faced similar waterborne lead regardless of  
 481 national origin.

482 Figure 4 displays earnings by father’s income quartile  
 483 for Pure Lead and Non-Lead towns, estimated using across-  
 484 town variation with the same town-level controls as the main  
 485 specification. Mixed Lead results are reported in the SI  
 486 Appendix. All coefficients are measured relative to a single  
 487 baseline group (Non-Lead  $\times$  High Leaching Risk  $\times$  Q1), so  
 488 both levels and gradients are directly comparable across town  
 489 types (SI Appendix, Section 5.4). Because lead-pipe cities  
 490 were wealthier and more industrial, the composition of father’s  
 491 income quartiles differs across town types: Q1 fathers account  
 492 for 17% of the Non-Lead  $\times$  High Leaching Risk population but  
 493 only 5% of the Pure Lead  $\times$  High Leaching Risk population.  
 494 Comparing gradients across town types in Panels (A) and  
 495

497 (B) must be interpreted with this compositional difference  
 498 in mind. Panel (C) avoids this concern by comparing within  
 499 Pure Lead towns across chemistry environments, where the  
 500 Q1 share is similar (5–6% in both leaching risk groups).

501 Panel (A) shows that in High Leaching Risk environments,  
 502 the two pipe types start at comparable Q1 earnings levels,  
 503 confirming that the baseline is similar after conditioning on  
 504 town characteristics. The Non-Lead gradient rises steeply: Q4  
 505 sons earn 10% more than Q1 sons ( $p < 0.01$ ). The Pure Lead  
 506 gradient is compressed: Q2 and Q3 sons show no earnings  
 507 advantage over Q1, and Q4 sons earn only 4% more ( $p < 0.05$ ).  
 508 This compression of the intergenerational earnings gradient  
 509 is consistent with a large literature documenting the role of  
 510 family background in shaping adult economic outcomes (33,  
 511 34). Panel (B) shows that in Low Leaching Risk environments,  
 512 both pipe types display upward-sloping gradients of similar  
 513 magnitude, confirming that lead infrastructure alone does  
 514 not compress the SES gradient. The compression is specific  
 515 to environments where chemistry promotes lead dissolution.

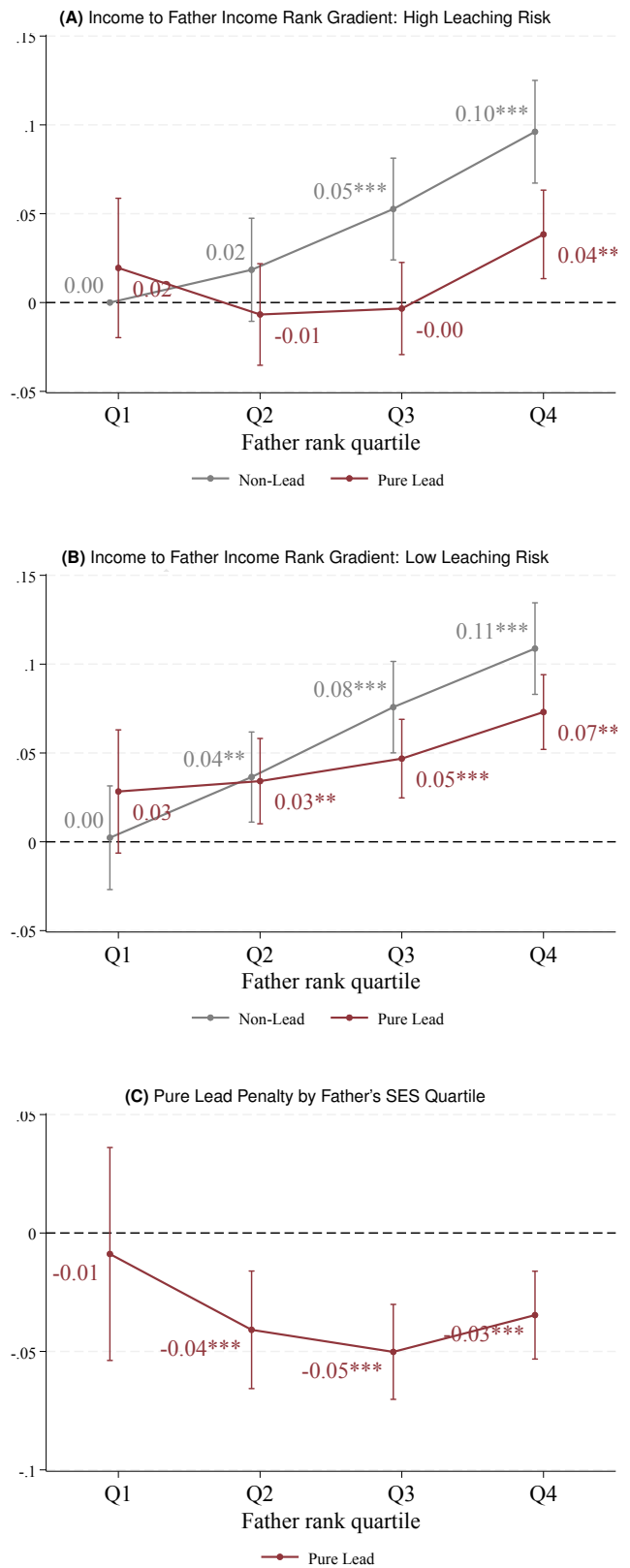
516 Panel (C) isolates the chemistry effect within Pure Lead  
 517 Lead towns by comparing High versus Low Leaching Risk  
 518 environments at each quartile. Corrosive water chemistry  
 519 reduces earnings at Q2 (−4%,  $p < 0.01$ ), Q3 (−5%,  $p < 0.01$ ),  
 520 and Q4 (−3%,  $p < 0.01$ ), with a small and insignificant  
 521 effect at Q1. The Non-Lead chemistry contrast is near zero  
 522 at every quartile (SI Appendix, Figure 9), confirming that  
 523 water chemistry affects earnings only in the presence of lead  
 524 pipes. Higher parental socioeconomic status did not shield  
 525 children from the labor market harm caused by lead exposure  
 526 in corrosive water environments: Q2, Q3, and Q4 sons all  
 527 face significant earnings penalties while Q1 sons do not.

## 528 Discussion

529 Our findings establish that childhood waterborne lead expo-  
 530 sure caused lasting labor market penalties. Men from lead  
 531 pipe towns with corrosive water earned less, worked fewer  
 532 weeks, and faced higher unemployment, with the within-pipe-  
 533 type chemistry contrast confirming that these effects are  
 534 driven by lead leaching rather than other correlates of urban  
 535 infrastructure. The penalty operates along multiple margins  
 536 of the labor market and is not offset by higher parental  
 537 socioeconomic status.

### 538 Mechanisms Underlying Labor Market Penalties.

539 The labor market penalties we document are not accompanied  
 540 by differences in self-reported years of completed education,  
 541 despite evidence from this same historical period that lead  
 542 exposure reduced cognitive test scores (10). The birth  
 543 cohorts in our sample (1890–1920) attended school during  
 544 the High School Movement, a period in which secondary  
 545 schooling expanded rapidly, driven by local institutional  
 546 investment, compulsory attendance laws, and community  
 547 wealth (20, 21). In this institutional environment, modest  
 548 cognitive disadvantages attributable to lead exposure may  
 549 not have translated into differences in reported attainment,  
 550 particularly given that self-reported schooling in Census data  
 551 is subject to heaping at grade-level milestones. The finding  
 552 that lead-exposed men earn less even within the same detailed  
 553 occupation is consistent with a broader literature linking  
 554 childhood lead exposure to reduced cognitive functioning  
 555 (11, 16, 35), as well as behavioral and conduct problems that  
 556 persist into adulthood (3, 29, 30). These capacities shape



557 Fig. 4. Earnings by father's income quartile and pipe type. All coefficients are relative  
 558 to Non-Lead × High LR × Q1, estimated using across-town variation with town-level  
 559 controls and IPW weights. Panel (A): High Leaching Risk environments. Panel (B):  
 560 Low Leaching Risk environments. Panel (C): Within-pipe-type chemistry contrast for  
 561 Pure Lead towns (High LR minus Low LR at each quartile); Non-Lead placebo near  
 562 zero at all quartiles (see SI Appendix). 90% confidence intervals. Standard errors  
 563 clustered at the pipe district level. \* $p < 0.10$ , \*\* $p < 0.05$ , \*\*\* $p < 0.01$ .  
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workplace performance and career progression in ways that Census data cannot directly observe, but that are consistent with the pattern of results we find.

**Magnitude of Effects and Comparison to Prior Research.** Our estimates are conservative for several reasons. First, our analysis captures only waterborne lead exposure. Men in our sample may also have been exposed through lead-based paint and occupational contact, meaning total lead exposure likely exceeded what our infrastructure-based measure reflects. Second, our intent-to-treat design assigns exposure based on town-level infrastructure and water chemistry rather than individual blood lead levels. Actual exposure varied within towns due to differences in water consumption, duration of residence, and household plumbing, so our estimates average over both exposed and unexposed individuals within treated towns.

A growing literature estimates the labor market effects of childhood lead exposure using modern policy variation. (36) find that reductions in airborne lead exposure are associated with a 3.5% increase in adult earnings for recent US cohorts. (6) estimate a 4% earnings effect from the phaseout of leaded gasoline in Sweden, though this is imputed from changes in educational attainment rather than directly observed wages. Other recent work has documented effects of lead exposure on birth outcomes (37), child development (4, 5, 38), and housing markets (39). Our estimates are broadly comparable in magnitude to these modern studies despite measuring a different exposure source (waterborne vs. airborne) in a different historical context, and despite the conservative features of our intent-to-treat design.

**Heterogeneous Effects and the Compression of Socioeconomic Advantages.** The compression of the intergenerational earnings gradient in lead-exposed towns speaks to a broader question about whether family resources can protect children from environmental harm. In our setting, where families could not selectively avoid exposure, higher parental socioeconomic status did not buffer children from the labor market consequences of lead. The within-pipe-type chemistry contrast confirms that this compression is driven by lead leaching rather than other features of lead-pipe cities: corrosive water chemistry penalizes Q2, Q3, and Q4 sons while leaving Q1 sons unaffected, and the same chemistry variation has no effect in Non-Lead towns. One interpretation is that lead exposure reduces the returns to the human capital investments that higher-SES families typically make in their children (33, 34). If lead impairs the cognitive and behavioral capacities needed to benefit from better schools, stronger networks, and greater parental investment, then the advantages these resources confer are diminished. This interpretation is consistent with the broader pattern of results: lead reduces earnings without reducing reported educational attainment, suggesting the harm operates through capacities that family resources cannot easily replace.

**Limitations.** Several limitations warrant consideration. First, we cannot directly observe individual cognitive ability, behavioral outcomes, or other valued worker attributes in the Census data. The within-occupation wage penalty is consistent with lead-induced reductions in these capacities, but we cannot distinguish this from other sources of within-occupation earnings variation such as firm quality, work intensity, or chronic health conditions caused by lead exposure

itself. Second, we observe each individual in a town at ages 0–10 but do not know the duration or precise timing of their exposure. The medical literature indicates that in-utero and early childhood exposure produce the largest developmental effects, but we cannot isolate these critical windows from later childhood exposure. Third, our analysis focuses exclusively on men because women’s labor market participation was limited in this period. The effects of lead exposure on women likely operate through other margins, including fertility, marriage, and household production, that we do not examine. Fourth, nonwhite men are nearly absent from lead-pipe cities in our linked sample, preventing us from examining racial heterogeneity in the effects of lead exposure. Given the prominent role of racial disparities in the modern lead literature (5), this remains an important question that our data cannot address. Fifth, while conditional differences in household head characteristics across pipe types are small after town-level controls (SI Appendix, Table 4), parents who chose to live in lead-pipe cities may have differed on dimensions we cannot observe. The within-pipe-type chemistry contrast mitigates this concern because it compares families within the same type of city, and water chemistry was geologically determined and not a factor in household location decisions. Sixth, our inverse probability weights account for mortality between an individual’s first census observation and 1940, but the direction of any remaining mortality selection is ambiguous: if the most severely affected individuals died before observation, our estimates understate the true effect, but surviving individuals may also carry chronic health burdens that contribute to the labor market penalties we measure.

**Implications for Policy and Future Research.** These findings are relevant for ongoing efforts to evaluate the costs and benefits of lead remediation. Current cost-benefit analyses of lead service line replacement focus primarily on health outcomes and cognitive gains. Our results indicate that the economic costs of lead exposure extend to labor market outcomes measured decades after childhood exposure, including reduced earnings, higher unemployment, and occupational downgrading. Accounting for these labor market costs alongside health effects provides a more complete picture of the returns to lead remediation, consistent with other work documenting broad economic returns to investments in water and sanitary infrastructure (40). We estimate annual per-person earnings losses of \$950 to \$1,350 in 2026 dollars, combining the wage penalty and increased unemployment (SI Appendix, Section 6.3). Over the 30-year working life observed in our data, cumulative losses range from \$28,500 to \$40,500 per exposed individual. The EPA estimates a central cost of \$4,700 per full lead service line replacement, with a range of \$1,200 to \$12,300 (32). The per-person earnings losses we estimate exceed even the high end of replacement costs by a factor of two to three, and these losses reflect only male earnings from a single exposure source, omitting health costs, mortality, and effects on women. The SES findings add a distributional dimension: lead exposure compresses the returns to family background, suggesting that cost-benefit frameworks focusing only on average treatment effects may miss important consequences for economic mobility. Future research should prioritize examining the effects of lead exposure on women, on racial and ethnic minorities

745 where data permit, and on outcomes later in the life course  
746 including health, disability, and retirement security (1).

747 **Materials and Methods**

749 **Data Sources.** The empirical analysis combines three primary  
750 data sources. First, individual-level data from the 1940 U.S.  
751 Decennial Census provide information on adult labor market  
752 outcomes (41). Second, we use the full-count 1900, 1910, and  
753 1920 U.S. Decennial Censuses with person identifiers (HistIDs)  
754 provided by IPUMS (41) and cross-decade linkages from the  
755 Census Tree (42, 43). Third, hand-digitized historical sources  
756 on municipal water infrastructure and chemistry come from *The*  
757 *Manual of American Water-Works*, which documents service pipe  
758 materials at the town level, and U.S. Geological Survey publications  
759 documenting historical measurements of pH and hardness (44, 45).  
760 Full details on data sources, digitization procedures, and linkage  
761 methodology and rates are provided in SI Appendix, Section 1.

762 **Sample Construction.** Our sample is males ages 20-50 in the  
763 1940 Census who can be linked to their childhood census record and  
764 who can be linked to town level water infrastructure and chemistry  
765 information. We restrict the analysis to men because women's  
766 labor market participation was limited in this period. We link  
767 each individual back to their childhood Census record when they  
768 were ages 0-10, so that each individual in the sample has an initial  
769 observation (in either 1900, 1910, or 1920) and a 1940 observation.  
770 We exclude individuals who cannot be linked back to a childhood  
771 Census record and those whose childhood town cannot be matched  
772 to infrastructure or water chemistry data. These sampling criteria  
773 yield 6.2 million men from 971 unique towns spanning all regions  
774 of the United States.

775 **Measuring Childhood Lead Exposure.** We categorize  
776 towns into three pipe material groups based on infrastructure  
777 records: Pure Lead (lead reported as the pipe material in use),  
778 Mixed Lead (a combination of lead and other materials reported),  
779 and Non-Lead (no lead pipes reported). To determine lead leaching  
780 potential, we predict theoretical lead solubility using Schock's  
781 solubility curves, which describe lead concentration as a non-linear  
782 function of pH and alkalinity (12, 27, 28). Because alkalinity  
783 is not consistently available in historical water quality records,  
784 we follow prior literature in using water hardness as a proxy and  
785 implement a Generalized Additive Model (GAM) fitted to digitized  
786 data from Schock's solubility nomograph. Towns are classified as  
787 High or Low Leaching Risk environments based on predicted lead  
788 solubility, with the threshold set at the 75th percentile of predicted  
789 values across all towns. This classification creates six exposure  
790 categories from the cross-classification of pipe type and leaching  
791 environment. As a robustness check, we provide the main results  
792 for two alternative classifications of High and Low Leaching Risk  
793 environments, at the 50th percentile of predicted lead values and at  
794 pH cutoffs used in prior literature (10, 11), in SI Appendix Section  
795 7. Full technical details on the predicted lead solubility model,  
796 including extrapolation strategies for extreme pH and hardness  
797 values, are provided in SI Appendix, Section 2.

798 We provide evidence in SI Appendix, Section 4, that conditional  
799 on the town-level fixed effects used in all specifications, remaining  
800 differences in pre-existing town characteristics across pipe types  
801 are small and statistically insignificant.

802 **Outcome Variables.** The primary outcome is wage and  
803 salary income in 1939, constructed from 1940 Census questions  
804 on annual earnings from wages and salaries (excluding self-  
805 employment income, business profits, investment income, and  
806 government transfers). We also examine weeks worked in the  
807 previous year, unemployment status, labor force participation, and  
808 occupational category. Occupations are classified using harmonized  
809 1950 occupation codes provided by IPUMS. We create two broad  
810 occupational tiers by splitting all occupations at the median of  
811 average occupation-level earnings in 1940: lower-tier occupations  
812 include laborers, service workers, and lower-skilled operatives, while

807 upper-tier occupations include professionals, managers, craftsmen,  
808 and skilled operatives. All monetary values are expressed in 1940  
809 dollars. Additional details on variable construction and treatment  
810 of missing or top-coded values are provided in SI Appendix,  
811 Section 3.

812 **Empirical Specification.** We estimate the effect of childhood  
813 waterborne lead exposure on adult labor market outcomes using  
814 ordinary least squares regressions of the form:

$$815 Y_{ic} = \beta_1 \text{LowLR}_c + \beta_2 \text{PureLead}_c \times \text{LowLR}_c \\ 816 + \beta_3 \text{PureLead}_c \times \text{HighLR}_c + \beta_4 \text{MixLead}_c \times \text{LowLR}_c \\ 817 + \beta_5 \text{MixLead}_c \times \text{HighLR}_c + X'_i \gamma + Z'_c \theta + \alpha_r + \varepsilon_{ic} \quad [1]$$

818 where  $Y_{ic}$  is the outcome for individual  $i$  who grew up in town  $c$ .  
819 The omitted category is Non-Lead pipes in High Leaching Risk  
820 (High LR) environments. The coefficient  $\beta_3$  captures the effect of  
821 Pure Lead pipes in High LR water, while  $\beta_1$  measures the main  
822 effect of Low LR environments for Non-Lead towns. Individual  
823 controls ( $X_i$ ) include age, race, household head's sex, occupation,  
824 and marital status, parental immigration status, homeownership,  
825 urban status, and group quarters status. Region fixed effects ( $\alpha_r$ )  
826 control for time-invariant regional differences based on childhood  
827 location. Town-level controls ( $Z_c$ ) are absorbed as categorical  
828 fixed effects for population, industry composition, metropolitan  
829 status, proximity to large cities, water system characteristics,  
830 and mining activity. Standard errors are clustered at the pipe  
831 district level, which groups towns sharing a common water system  
832 (824 clusters). Inverse probability weights are used to weight an  
833 individual back to their representation in their original census  
834 observation. Unweighted results are comparable and reported in  
835 SI Appendix 7.

836 For occupational sorting analysis, we estimate equation (1)  
837 separately for an indicator of lower-tier occupation. For within-  
838 occupation wage analysis, we augment equation (1) with 223  
839 detailed occupation fixed effects, comparing earnings among  
840 workers in the same occupation. For heterogeneity analysis by  
841 childhood socioeconomic status, we interact equation (1) with  
842 the four father income quartiles. Complete variable definitions,  
843 regression specifications, and robustness checks are provided in SI  
844 Appendix, Sections 3, 5, and 7.

845 **Validation of Research Design.** To support the validity  
846 of our research design, we conduct several tests. First, we  
847 show that conditional on the town-level fixed effects used in all  
848 specifications, towns with different pipe materials are balanced on  
849 pre-exposure demographic and economic characteristics observed  
850 in the 1900 Census (SI Appendix, Table 3), and that household  
851 head characteristics are similarly balanced at the individual level  
852 (SI Appendix, Table 4). Second, we demonstrate that leaching  
853 risk is not systematically correlated with pipe type choice (SI  
854 Appendix, Figure 8). Third, we show that linkage rates from  
855 childhood to adulthood do not vary systematically with pipe type  
856 or leaching environment (SI Appendix, Figure 2). Finally, we  
857 conduct robustness checks including alternative chemistry cutoffs  
858 and additional specification tests (SI Appendix, Section 7).

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